



Integrated population model of Antipodean albatross for simulating management scenarios

Technical report prepared for Department of Conservation – June 2021

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Cover Notes

To be cited as:

Richard, Yvan (2021). Integrated population model of Antipodean albatross for simulating management scenarios, 37 pages. Technical report prepared for Department of Conservation – June 2021.

Cover image:



<https://www.flickr.com/photos/angrysunbird/2549595283>

EXECUTIVE SUMMARY

2 Antipodean albatross *Diomedea antipodensis antipodensis* is endemic to New Zealand,
with the quasi-totality of the population nesting on Antipodes Island. The species is
4 classified as Nationally Critical due to a potential demographic decline. Threats to the
population include incidental mortality in fisheries, climate change, and depredation by
6 exotic species.

The objective of this project was to provide a tool that allows stakeholders to explore the
8 potential impact of threats and the demographic outcomes of management strategies.
Using the tool, simulations of the demographic impact of different scenarios may be
10 carried out so that management strategies can be assessed and prioritised.

A small subset of the population of Antipodean albatross has been studied since 1994, and
12 these field data were used to perform the simulations. A Bayesian integrated population
model was developed to estimate the main demographic parameters of the population.
14 The model considered detectability of individuals, inter-annual variability, movements
in and out of the study area, and data censoring; it was fitted using the software Stan.

16 The model results indicated that the probability of detecting individuals decreased from
2007 onwards, but this finding did not explain the observed decline in adult annual
18 survival. The estimated annual survival rate for females was estimated to decline from
0.947 (95% c.i.: 0.914 – 0.974) in the period from 1994 to 2004, to 0.882 (95% c.i.: 0.814 –
20 0.94) after 2005. Estimated survival for males was higher, at 0.946 (95% c.i.: 0.913 – 0.972)
and 0.927 (95% c.i.: 0.887 – 0.961) for the two periods. Breeding success also declined
22 between the two periods, from 72.4% (95% c.i.: 65.8% – 78.6%) from 1994 to 2004 to 63.7%
(95% c.i.: 53.4% – 73%) subsequently.

24 Under the current scenario, simulations suggest a significant decline of the population,
with an annual growth rate of -4.84% (95% c.i.: -6.07% – -3.65%). Limitations in the data
26 and in the model assumptions may cause the decline to be overestimated; however, the
results raise concerns about the sustainability of the population.

28 The simulation tool is aimed to assist conservation managers with the prioritisation of

management strategies to mitigate threats to the Antipodean albatross population and to
30 guarantee the persistence of this species.

DRAFT - Not to be quoted

1. INTRODUCTION

32 The seabird species Antipodean albatross (*Diomedea antipodensis antipodensis*) is endemic
to New Zealand and consists of two subspecies, Antipodean albatross (*D. a. antipodensis*)
34 and Gibson's albatross (*D. a. gibsoni*). The subspecies Antipodean albatross breeds
almost exclusively on Antipodes Island, with a few pairs breeding on Chatham and
36 Campbell islands, whereas Gibson's albatross breeds on Auckland Island. The species is
classified as Endangered by the International Union for Conservation of Nature (BirdLife
38 International 2018), and each subspecies is classified individually as Nationally Critical
in New Zealand (Robertson et al. 2017).

40 The population of Antipodean albatross is exposed to a number of threats, at sea and on
land. They are caught incidentally in surface-longline fisheries in New Zealand waters
42 and globally (Richard & Abraham 2017). Chicks used to be depredated by mice at the
nest, although mice have been eradicated from Antipodes Island since 2016. Climate
44 change may also impact the population indirectly, increasing heat stress to chicks and
affecting the distribution or abundance of prey species.

46 On Antipodes Island, a 29-ha (0.29-square kilometre) area of the Antipodean albatross
population has been monitored every year since 1994, except in 2006. Field data from this
48 area (Elliott & Walker 2020) and quantitative modelling (Edwards et al. 2017) suggest a
population decline since 2007, via a decline in female survival and in breeding success,
50 and an increase in recruitment age. Tracking data of individual at-sea movements also
suggest a potential change in the foraging grounds over time (Elliott & Walker 2020).

52 Tracking at-sea movements also allowed the identification of fisheries with the highest
overlap with the species (Bose & Debski 2020). A number of mitigation techniques exist
54 to reduce the level of incidental captures in fisheries and are already in place in a number
of fisheries, in New Zealand and worldwide (Løkkeborg 2011).

56 The main objective of this project was to develop an online tool to facilitate the
prioritisation of management strategies around population threats. The online tool
58 allows the running of simulations of the fate of the population under different scenarios,

leading to the identification of strategies with the highest positive impact on the
60 population. The simulations rely on estimates of the main demographic parameters of
this subspecies. A Bayesian integrated population model was developed for this purpose,
62 based on the individual capture-recapture data that have been collected in the study area
on Antipodes Island since 1994.

64 2. METHODS

The Antipodean albatross subspecies breeds almost exclusively on Antipodes Island
66 (Agreement on the Conservation of Albatrosses and Petrels 2009). When breeding, a
single egg is laid on a nest consisting of a low pedestal build of soil and vegetation, often
68 re-used between breeding attempts. It takes a year for an egg to produce a fledgling. For
this reason, adults can only breed every second year when successful. Fledglings spend
70 the first few years at sea before returning to the colony, and subsequently spend another
year or more before breeding for the first time.

72 Since 1994, a 29-ha (0.29-square kilometre) area on Antipodes Island has been surveyed
every year, except in 2006; the most recent survey was in 2021. Survey visits to the island
74 were generally conducted in January, so that the outcome of the previous year's breeding
attempts could be observed, and new breeding attempts could also be recorded. Each
76 visit was on average for a month to allow sufficient time to survey the birds present and
to band any new birds in the study area. Due to the remoteness of the island and its
78 limited accessibility, logistic constraints led to variation in the exact timing and length of
visits between years.

80 The data collected in the field consist of the date and location of detected banded
individuals at the site, their breeding status and stage, and their sex when identifiable.

82 Additionally, a buffer around the study area was frequently visited, in addition to two
other blocks on the island. In these areas, the sightings and breeding status of banded
84 individuals were also recorded, and identified as being outside the study area. A
description of the field data is presented in Edwards et al. 2017

86 The data were aggregated to create individually- and annually-based capture histories,
representing the state of individuals each year between 1994 and 2021. Individuals were
88 categorised into three age classes: juvenile (between fledging and first return to the
colony), pre-breeder (from first return to first breeding at the colony), and adult (after
90 first breeding). Eight observed states were represented:

1. adult breeding inside the study area;
- 92 2. adult non-breeding inside the study area;
3. adult outside the study area (breeding or not);
- 94 4. pre-breeder inside the study area;
5. pre-breeder outside the study area;
- 96 6. juvenile;
7. dead;
- 98 8. not seen.

Adults sighted both inside and outside the study area one year were considered inside the
100 study area. Adults only sighted outside the study area were not split between breeders
and non-breeders as their breeding status cannot be identified precisely (especially for
102 birds seen early in the season). Because surveys of the study area overlapped between
the end of the previous breeding season and the beginning of the next one, the aggregated
104 data were prepared to represent the status of the population just before breeding occurs;
i.e., chicks of the current breeding year first appear in the prepared data the following year
106 after fledging (if successful). Only birds banded within the study area were included in
the final dataset.

108 Nest success was recorded at the nest level, as the nesting individuals might not
necessarily be seen, and nests were considered successful if they produced a fledgling.
110 A successful nests could either have a chick being very close to fledging at the last
observation, or empty but showing indications of recent breeding activity without
112 showing any sign of failure (e.g.; broken shells, dead body parts).

2.1 Integrated population model

114 To estimate the main demographic parameters of the population of Antipodean albatross,
a multi-state Bayesian capture-recapture model was developed. This type of model aims
116 to alleviate the main biases in the data, which are common to most population survey
data.

118 The state of an individual can be unknown, and an individual may be undetected but
still alive. Individuals may be undetected in a given year for several reasons. They could
120 be at sea, such as juveniles, adults previously breeding successfully or on a “sabbatical”
year, or breeding adults on a foraging trip may not be detected during short visits to the
122 island. Undetected individuals could also be present at the colony, but outside the study
area.

124 For these reasons, the “actual” state of individuals was considered as a latent variable
in the model, with year-to-year transitions between the states determined by explicit
126 biological rules. For example, an adult cannot become a juvenile, or an adult breeding
successfully cannot breed again the following year. In addition, an observation process
128 was considered, linking the latent state to the observed state, and determined by both the
survey effort and the birds’ behaviour.

130 2.1.1 Latent states

A total of eight latent states were considered in the model, different from the observed
132 states:

1. adult breeding inside the study area;
- 134 2. adult breeding outside the study area;
3. adult non-breeding inside the study area;
- 136 4. adult non-breeding outside the study area;
5. pre-breeder inside the study area;

138 6. pre-breeder outside the study area;

7. juvenile;

140 8. dead.

The transition matrix between the eight latent states required specifying the probability of
142 being in each latent state given the previous one, representing 64 transition probabilities.

For juveniles (J), pre-breeders (PB), breeding adults (B), and non-breeding adults (NB),
144 the probabilities of changing to a different live state given the previous state were:

$$P(\text{PB}_t | \text{J}_{t-1}) = R_a \phi_J, \quad (1)$$

$$P(\text{B}_t | \text{PB}_{t-1}) = B_a \phi_{\text{PB}}, \quad (2)$$

$$P(\text{B}_t | \text{NB}_{t-1}) = P(\text{breed} | \text{non-breeder}) \phi_s, \quad (3)$$

$$P(\text{NB}_t | \text{B}_{t-1}) = \begin{cases} 1 & \text{after a successful breeding attempt,} \\ (1 - P(\text{breed} | \text{fail})) \phi_s & \text{after a failed breeding attempt,} \end{cases} \quad (4)$$

where t is the year, $\phi_{\{\text{J}, \text{PB}, \text{s}\}}$ the annual survival rate of juveniles, pre-breeders, and adults
146 of sex s , respectively, R_a the probability of a juvenile of age a returning to the colony,
 B_a the probability of a pre-breeder of age a breeding for the first time, $P(\text{breed} | \text{fail})$
148 the probability of an adult breeding in a particular year, given it was an unsuccessful
breeder the previous year, $P(\text{breed} | \text{non-breeder})$ the probability of an adult breeding in
150 a particular year, given it was a non-breeding adult the previous year.

When the sex was unknown, conditional probabilities were used; e.g., the annual survival
152 rate of an individual of unknown sex was $P(\text{♀})\phi_{\text{♀}} + (1 - P(\text{♀}))\phi_{\text{♂}}$, where $P(\text{♀})$ is the
probability that an individual in the study area is a female.

The probabilities of remaining in the same live state from one year to the next were:

$$P(J_t|J_{t-1}) = (1 - R_a)\phi_J, \quad (5)$$

$$P(PB_t|PB_{t-1}) = (1 - B_a)\phi_{PB}, \quad (6)$$

$$P(NB_t|NB_{t-1}) = (1 - P(\text{breed}|\text{non-breeder}))\phi_s, \quad (7)$$

$$P(B_t|B_{t-1}) = \begin{cases} 0 & \text{after a successful breeding attempt,} \\ P(\text{breed}|\text{fail})\phi_s & \text{after a failed breeding attempt,} \\ (1 - P(\text{success}))P(\text{breed}|\text{fail})\phi_s & \text{after an unknown outcome.} \end{cases} \quad (8)$$

In addition, the transition probabilities were multiplied by the probability of moving inside or outside the study area, depending on the state:

$$P(\text{Out}_t|\text{In}_{t-1}) = E_s, \quad (9)$$

$$P(\text{In}_t|\text{Out}_{t-1}) = I_s, \quad (10)$$

$$P(\text{Out}_t|\text{Out}_{t-1}) = 1 - I_s, \quad (11)$$

$$P(\text{In}_t|\text{In}_{t-1}) = 1 - E_s, \quad (12)$$

154 where E_s is the probability of an individual of sex s moving out of the study area
 (emigrate), and I_s the probability of an individual of sex s moving into the study area
 156 (immigrate).

The probabilities of being dead (D) in a particular year were:

$$P(D_t|J_{t-1}) = 1 - \phi_J, \quad (13)$$

$$P(D_t|PB_{t-1}) = 1 - \phi_{PB}, \quad (14)$$

$$P(D_t|B_{t-1}) = 1 - \phi_s, \quad (15)$$

$$P(D_t|NB_{t-1}) = 1 - \phi_s, \quad (16)$$

$$P(D_t|D_{t-1}) = 1. \quad (17)$$

The probability of impossible transitions—e.g., from adult to juvenile or to pre-breeder,
158 from pre-breeder to juvenile, and from dead to alive—were fixed to zero.

The adult annual survival rate was estimated independently for females and males, and
was allowed to vary randomly between years, with the survival rate $\phi_{s,t}$ for sex s at year
 t being defined on the logit scale as:

$$\text{logit}(\phi_{s,t}) = \text{logit}(\bar{\phi}_s) + \epsilon_{s,t}s_s, \quad (18)$$

where $\bar{\phi}_s$ is the mean survival rate across years for sex s , $\epsilon_{s,t}$ is the normally-distributed
160 random effect for each sex and year, and s_s is the sex-specific variability of the random
effect among years.

162 The annual survival rate of juveniles and pre-breeders was assumed to be constant over
time, and the same between males and females in the model.

164 As for adult survival, breeding success, i.e., the probability that a nest produces a
fledgling, was also modelled as a random effect over time.

166 The probability R_a of a juvenile of age a returning to the colony and becoming a pre-
breeder was set to 0 at ages below the minimum observed age at first return (3 years),
168 and set to 1 for birds of age 9 and above, as all birds are expected to have returned to the
colony by age 9 (G. Elliott, pers. comm.). The age-specific probability of return for birds
170 aged 3 to 8 was modelled as a random effect.

Similarly, the probability B_a of a pre-breeder of age a to become a breeder for the first
172 time was set to 0 for birds under 7 years old, the minimum recorded breeding age. The
age-specific probability of first breeding for birds aged 7 to 20 was modelled as a random
174 effect. The probability for birds aged 21 and above was set to be constant to represent
the long tail in the distribution of age at first breeding (i.e., some birds take a long time
176 to breed or do not breed)

Both R_a and B_a were dependent on age, but assumed not to vary with year.

178 2.1.2 Observation process

In the model, latent states are related to observed states via an observation matrix,
180 representing the probability of recording any of the eight observed states given a latent
state (one of 8 latent states, different from the observed states).

182 The probability of detection was estimated separately in the model for:

- breeding adults inside the study area,
- 184 • non-breeding adults inside the study area that previously bred successfully,
- other non-breeding adults inside the study area,
- 186 • pre-breeders inside the study area,
- adults and pre-breeders outside the study area,
- 188 • juveniles (outside the study area by definition),
- dead individuals.

190 There were only a few recorded observations of juveniles and dead individuals, with
all juveniles and most deaths being recorded at sea. For this reason, their detection
192 probability was assumed to be constant among years.

Because year-to-year variations are most likely to reflect the timing and amount of
observations on the island, the other detection probabilities were allowed to vary among
years, but with the same annual variability among them; they were defined as:

$$\text{logit}(\gamma_{x,t}) = \text{logit}(\gamma_x) + \epsilon_t s, \quad (19)$$

194 where $\gamma_{x,t}$ is the detection probability of birds of category x at year t , $\text{logit}(\gamma_x)$ the average
detection probability for category x among years, ϵ_t the random annual effect of year t
for all categories, and s the variability among years for all categories.

196 For 2006, when the population was not surveyed, all detection probabilities were fixed to
zero.

198 **2.1.3 Model fitting**

The model was written in the Stan language and fitted in the R statistical package (R Core
200 Team 2019) using the *rstan* library (Stan Development Team 2020).

Stan was chosen over alternatives such as Bugs or JAGS as it implements the no-U-
202 turn sampler (NUTS; Hoffman & Gelman 2014) which improves model convergence and
allows fitting times to be reduced by an order of magnitude (from days to hours).

204 One disadvantage of Stan is that it does not support the direct sampling of discrete
parameters. Nevertheless, multi-state models can still be fitted by marginalising discrete
206 latent states, i.e., summing at each time step the likelihood of the observed state over
all possible latent states, iteratively over each individual capture history (Yackulic et al.
208 2020).

The model was fitted using Markov chain Monte Carlo (MCMC) methods, using four
210 chains, for 6,000 iterations, after a burn-in period of 5,000 iterations.

The code of the Stan model is provided in Appendix A.

212 **2.2 Population simulations**

The main aim of this project was to provide stakeholders with a tool to simulate the fate
214 of the Antipodean albatross population under different scenarios. For this purpose, an
interactive online application written in R and using the Shiny framework was developed.

Because the demographic model does not provide the latent state of individuals at
each time step directly due to the marginalisation of discrete latent variables, the initial
population structure for the simulations was derived separately. For this purpose, the
latent state at each time step for each individual was drawn randomly from the previous
state and the observed state. Using Bayes' theorem, the probability of an individual to be
in the latent state Π_i given the observed state O is:

$$P(\Pi_i|O) = \frac{P(O|\Pi_i)P(\Pi_i)}{P(O)}, \quad (20)$$

216 where $P(O|\Pi_i)$ is the probability of the observed state O given the latent state, which is
the detection probability of that state, as estimated by the model. $P(\Pi_i)$ is the probability
218 of state Π_i and is the transition probability from the previous latent state, as estimated by
the model. $P(O)$ is the probability of the observed state, and is the sum of observing O
220 given all possible latent states, i.e., $\sum_k P(O|\Pi_k)P(\Pi_k)$. In addition, the probability of a
dead individual at a given time step was set to zero when the individual was subsequently
222 detected alive. The process was repeated for each of the 6 000 MCMC samples from the
model, and the resulting population structure in 2021 – and its uncertainty – was taken
224 as the initial population for the simulations. Pre-breeders and adults outside the study
area were not included, to simulate only the population inside the study area and the
226 juveniles that fledged from there.

The population size from the simulations was scaled up by the ratio of the total number
228 of breeding pairs on the island to the number of breeding pairs inside the study area. The
total number of breeding pairs was estimated from extensive surveys of the whole island
230 in 1994, 1995, and 1996. The scaling of the studied population size to the whole island,
therefore, assumes that the ratio did not change over time. The proportion of the number
232 of breeding pairs that were inside the study area was estimated to be 2.7332% averaged
across the three censuses (Elliott & Walker 2020), and the inverse of this value (36.58715)
234 was used to scale up the simulation population size to the whole island.

The population simulations consisted of predicting the fate of each individual in the
236 initial 2021 population, and of new fledglings produced each year, every year for 30
years, based on the demographic parameters estimated in the model. For each simulated
238 year, an actual year between 2008 and 2020 was first drawn randomly to represent
the interannual variability estimated in the model, while considering only the most
240 recent years. The drawn year defined the value of survival rates and breeding success.
Surviving individuals were drawn following a Bernoulli process with a probability equal
242 to the survival rate of the drawn year and of the individual class (juvenile, pre-breeder,
adult female, or adult male). Juveniles and pre-breeders either remained in their age
244 class or moved to the next one depending on the age-specific transition probabilities.
Adults breeding that year were then drawn according to the probability of breeding,

246 depending on whether they bred successfully (or not) the previous year. The success
of breeding adults was then drawn randomly from the probability of success of that
248 year. Among successful breeders, the number of fledglings produced was taken as the
minimum number of female or male adults, and new individuals of age 0 were created,
250 with a sex assigned randomly with a probability of 0.5. This process was then repeated
iteratively for the 30 simulated years, and for each iteration of the MCMC methods.

252 In the online tool, scenarios are specified in terms of direct impacts, affecting specific
demographic parameters. Threats can impact the annual survival rate of juveniles,
254 pre-breeders, adult males, and adult females separately, or can also impact breeding
probability or breeding success. The threats can be defined as being either already
256 present, in which case the impact is removed from the population in the simulations, or
potential, with the impact added to the population. For example, to assess the potential
258 effect of introducing new mitigation measures in fisheries, the impact would need to be
specified as already present, and the incidental mortalities would be removed from the
260 population in the simulations.

Impacts may be specified as an absolute change in the demographic parameter, or as
a number of individuals for survival rates. When using individuals, the impact is
converted to the absolute change in survival rate, Δ , based on the total number of
individuals in the affected category:

$$\Delta = S' - S = 1 - \frac{(1 - \Phi)N - I}{N} - \Phi, \quad (21)$$

where S' is the new survival rate, Φ the survival rate of the population category (juvenile,
262 pre-breeder, adult female, or adult male), N the scaled-up number of individuals in the
category, and I the number of mortalities caused by the threat. The conversion of impacts
264 from individuals to a change in demographic rates assumes that the impact of threats is
consistently proportional to the population size.

266 Multiple threats and impacts may be specified for a given scenario. In that case,
the overall change in demographic parameters is calculated by summing the absolute
268 changes across threats and impacts within each demographic parameter.

Upon completion of the simulations, the mean and 95% credible interval of the
270 population size, of the number of annual breeding pairs, and population mean annual
growth rate, and the mean population structure are calculated and reported, in tables
272 and figures.

For illustration purposes, two hypothetical scenarios were simulate here, representing
274 two existing threats; each threat resulted in the death of 500 individuals, but only of
juveniles in one scenario, and only of adults in the other scenario (male and female).

276 **3. RESULTS**

3.1 Model parameters

278 The MCMC traces indicated that the model converged reasonably well, as the four chains
were well mixed and did not show significant autocorrelation (see Appendix B for the
280 MCMC traces and values of each demographic parameter estimated by the model). One
exception was the parameter related to the detection probability, which converged but
282 showed marked autocorrelation. (This autocorrelation will be corrected during a longer
and thinned fitting of the final model, upon finalising this report.)

284 The estimated adult annual survival rate between 1994 and 2020 showed changes over
time (Figure 1). Before 2005, the estimated survival rate was similar between sexes, with
286 an annual mean of 0.947 (95% c.i.: 0.914 – 0.973). From 2005, however, estimated female
survival declined to a mean of 0.882 (95% c.i.: 0.814 – 0.94); female survival was lowest in
288 2013, estimated at 0.821 (95% c.i.: 0.752 – 0.883). In contrast, male survival only slightly
declined to a mean of 0.927 (95% c.i.: 0.887 – 0.961), with a minimum around 0.90 in 2007.

290 The estimated survival in the the three most recent years (2018 to 2020) suggested a
possible increase to levels similar to estimates before 2004, with female adult survival
292 reaching 0.929 (95% c.i.: 0.861 – 0.976) in 2020, and adult male survival at 0.971 (95% c.i.:
0.943 – 0.991).

294 The annual survival rate of juveniles and pre-breeders, assumed to be constant among

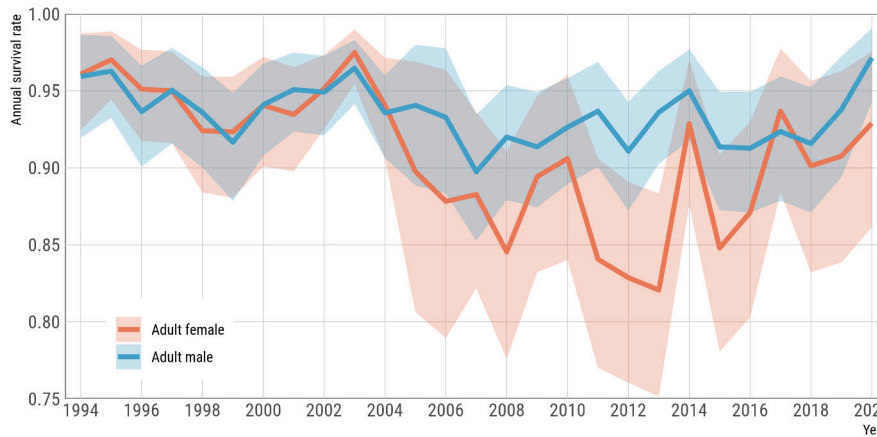


Figure 1: Adult annual survival for female and male Antipodean albatross between 1994 and 2020, estimated from the demographic model. Lines indicate the mean, shading the 95% credible interval.

years, was estimated at 0.879 (95% c.i.: 0.869 – 0.888) and 0.922 (95% c.i.: 0.913 – 0.931),
 296 respectively.

The change of adult survival rates over time was significant even though the interannual
 298 variability in the probability of detection was controlled in the model. The detection
 probability also showed a decrease over time, i.e., after 2006 (Figure 2).

300 despite controlling in the model for the inter-annual variability in the probability of
 detection, which also showed an overall decrease after 2006 (Figure 2).

302 The interannual change in detectability, applied to all individual types present on the
 island, was related to both the timing and length of the field seasons on the island
 304 (Figure 3). Estimates of detectability were highest when the field season started early
 (early December) and when the survey effort was high, both in the number of days
 306 with recorded field observations, and in the total number of recorded observations in
 the season.

308 Amongst the years with the lowest detectability, 1995 and 2020 were characterised by a
 low number of field days and observations, and started late in the season (mid-February
 310 and mid-March, respectively). In contrast, the highest estimated detectability was in
 2003, when the field season was both the second earliest (mid-December) and the second
 312 longest (60 days of observations).

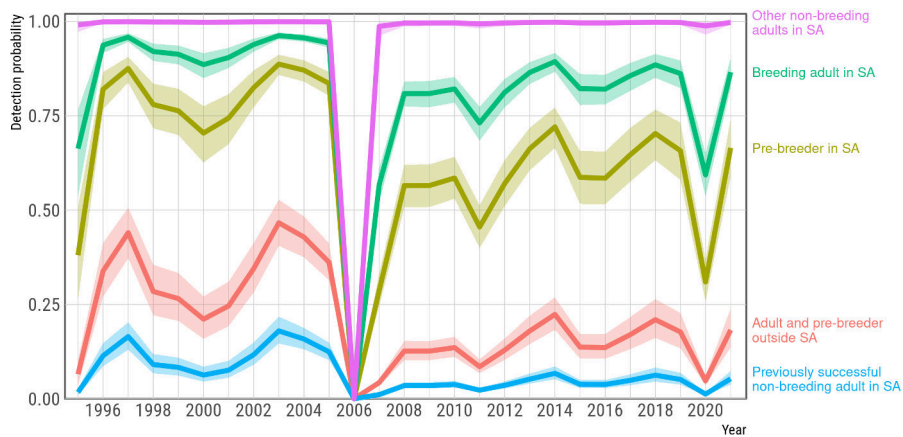


Figure 2: Detection probability of individuals inside the study area (SA) for breeding adults, non-breeders that were previously successful breeders, other non-breeders, and pre-breeders, and for adults and pre-breeders combined outside the study area. Lines indicate the mean, shading the 95% credible interval.

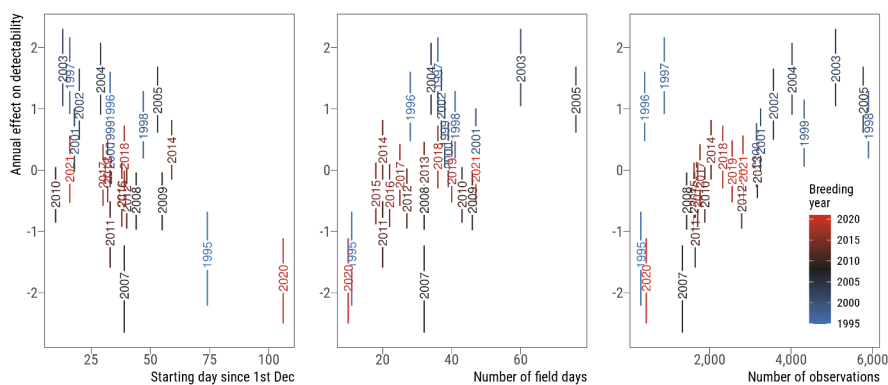


Figure 3: Relation between the interannual variability of the probability of detection and the timing and effort of population surveys. The timing of surveys was measured here as the number of days between the 1 December preceding the breeding season and the first day of recorded observations. Observation effort is in the number of days with observations, and the total number of observations recorded during the breeding season. The annual effect on detectability is shown as the 95% credible interval of the annual random effect as estimated in the model, and the label showing the year of the field season is centred on the mean estimate.

The estimated probability of detection varied significantly between the types of individuals considered in the model (Table 1 and Figure 2). This probability was around 5.2% for non-breeding adults that were successful breeders the previous years, 18.0% for adults and pre-breeders outside the study area, 66.1% for pre-breeders inside the study area, 86.4% for adults breeding inside the study area, and 99.7% for non-breeding adults

318 that were not successful breeders in the previous year. Additionally, the detectability was
 estimated close to zero for both juveniles and dead individuals, with a mean of 0.019%
 320 (95% c.i.: 0% – 0.073%) and 0.083% (95% c.i.: 0.054% – 0.118%), respectively.

Table 1: Mean estimates (and credible interval, c.i.) of the probability of detection among the different individual types in the Antipodean albatross population considered in the demographic model (SA, study area).

Type	Mean	95% c.i.
Breeding adult in SA	0.864	0.816 – 0.900
Previously successful non-breeding adult in SA	0.052	0.036 – 0.072
Other non-breeding adults in SA	0.997	0.992 – 1.000
Pre-breeder in SA	0.661	0.575 – 0.736
Adult and pre-breeder outside SA	0.180	0.132 – 0.234

The probability of breeding was estimated in the model, and assumed to be constant
 322 among years. For adults that were failed breeders the previous year, the probability
 of breeding was estimated at 70.5% (95% c.i.: 68.6% – 72.3%). The probability was
 324 significantly lower for other individuals that were previously non-breeders, at 64.1%
 (95% c.i.: 62.8% – 65.4%). For adults that were successful breeders in the previous year,
 326 this probability was zero.

As for survival, breeding success was also allowed to vary among years in the model.
 328 Modelled as the probability that a nest successfully produces a fledgling, breeding
 success also declined between the period 1994–2004 and 2005–2021 (Figure 4). Prior to
 330 2005, the mean breeding success was estimated at 72.4% (95% c.i.: 65.8% – 78.6%), but at
 63.7% (95% c.i.: 53.4% – 73%) after 2005.

332 To take into account bird movements in and out of the study area for the estimation
 of survival rates, the probability of individuals that were inside the study area leaving
 334 the area, and conversely the probability of individuals that were outside the study area
 returning to it, were estimated for females and males independently, and assumed to be
 336 constant among years. These probabilities suggest that females are less faithful to their
 area than males, as females had a 9% (95% c.i.: 8.1% – 10%) probability of leaving the
 338 study area, compared with 4% (95% c.i.: 3.5% – 4.6%) for males. Similarly, females had
 an estimated probability of 17.7% (95% c.i.: 15.2% – 20.3%) to return to the colony after

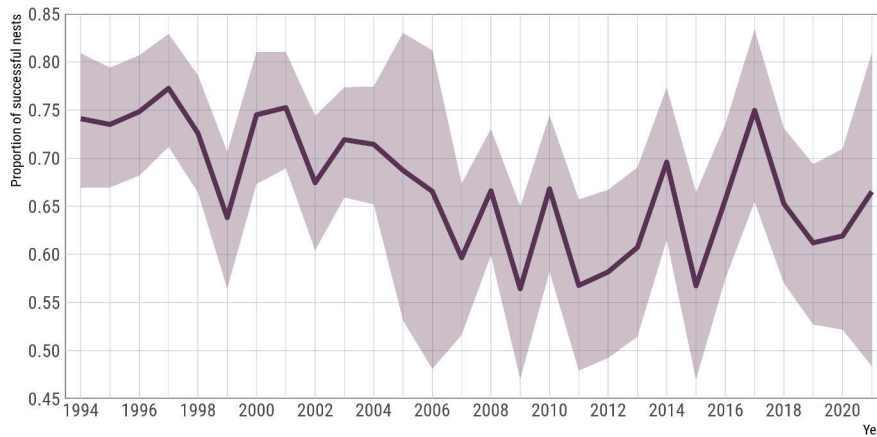


Figure 4: Breeding success by year for Antipodean albatross between 1994 and 2021, measured as the proportion of nests producing a fledgling. Line indicates the mean, shading the 95% credible interval.

340 leaving it, compared with 25.4% (95% c.i.: 21.9% – 29.1%) for males.

The ages at first return and at first breeding were also estimated in the model (Figure 5).

342 The age at first return varied between 3 and 9 years, with an average at 6.26 years. The
 344 minimum age at first breeding was 7 years, and by age 13, half of the individuals had
 bred at least once, although some individuals did not breed at all.

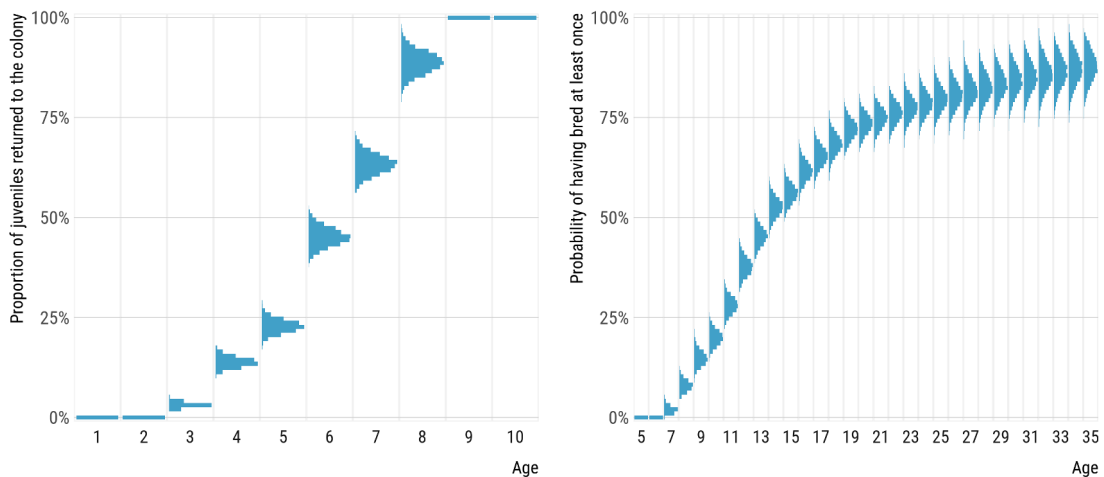


Figure 5: Proportion of individuals that returned to the colony (left) and proportion of individuals that bred at least once as function of age (right). For each age, a histogram of the Markov chain Monte Carlo values is shown as estimated by the model.

3.2 Online simulation tool

346 Based on the demographic parameters obtained from the model, an online application
was developed to simulate the population dynamics of Antipodean albatross under
348 different scenarios (see a screenshot of the online simulation tool in Figure 6).

The structure of the population in 2021 was used for the initialisation of the simulations,
350 and was obtained from drawing iteratively the latent state of each individual in the study
area each year when the state was unknown (examples of the predictions of individual
352 state are shown in Figure 7).

The number of number of breeding pairs inside the study area from on-site surveys
354 was similar to the estimate derived from the model estimates (Figure 8). Nevertheless,
the model estimate was higher overall. This difference was due to the model estimate
356 including the individuals that are not detected during surveys.

The population in 2021 used to initialise the simulations was estimated inside the study
358 area at 90 (95% c.i.: 81 – 100) breeding pairs, and 762 (95% c.i.: 726 – 801) total individuals.
Scaling up to the entire island, these estimates represent a total of 3,292 (95% c.i.: 2,964 –
360 3,659) breeding pairs and 27,893 (95% c.i.: 26,562 – 29,306) total individuals.

On average, the population consisted of 15.7% juvenile, 21.3% pre-breeders, 37.5% non-
362 breeding adults, 17% successful breeding adults, and 8.6% unsuccessful breeding adults.

In the current context, i.e., without specifying any management scenario, simulations
364 predicted a population decline of 4.84% (95% c.i.: 3.65% – 6.07%) with the total annual
number of breeding pairs in the study area decreasing from 90 (95% c.i.: 81 – 100) to 11
366 (95% c.i.: 4 – 21) after 30 years (“Current context” in Figure 9). Scaling up the study area
population to the entire island, this estimate corresponded to a decline from 3,292 (95%
368 c.i.: 2,964 – 3,659) breeding pairs to 401 (95% c.i.: 146 – 768), or for the whole population,
from 27,893 (95% c.i.: 26,562 – 29,306) birds to 6,412 (95% c.i.: 4,244 – 9,183).

370 When simulating a hypothetical scenario of mitigating an existing threat causing the
death of 500 juveniles, the rate of decline decreased to 3.3% (95% c.i.: 2.1% – 4.6%); when
372 the mortalities only affected adults, the rate further decreased to 2.7% (95% c.i.: 1.5% –

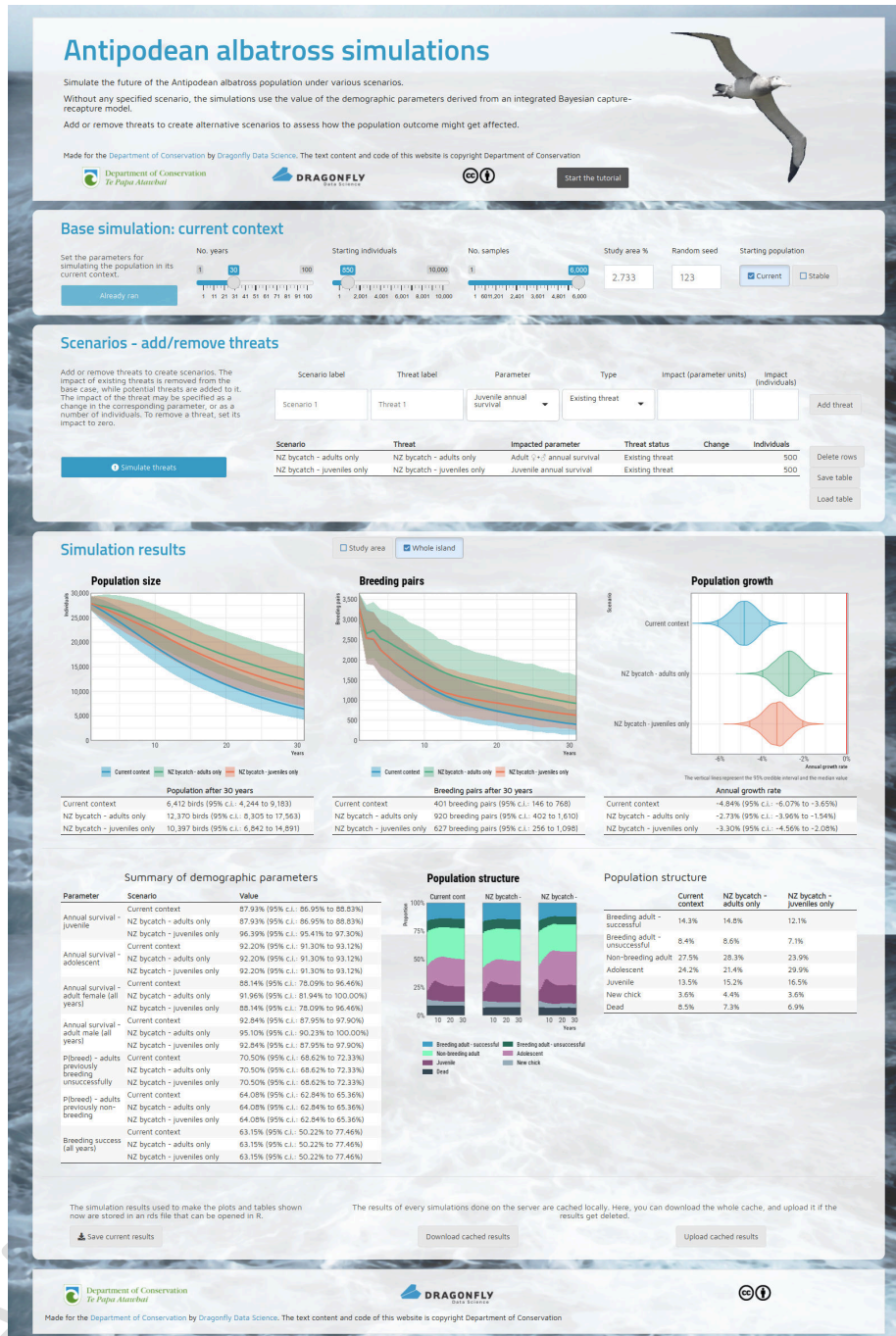


Figure 6: Screenshot of the online application tool to run predictions of the Antipodean albatross population in the future under different scenarios.

4%) (Figure 9).

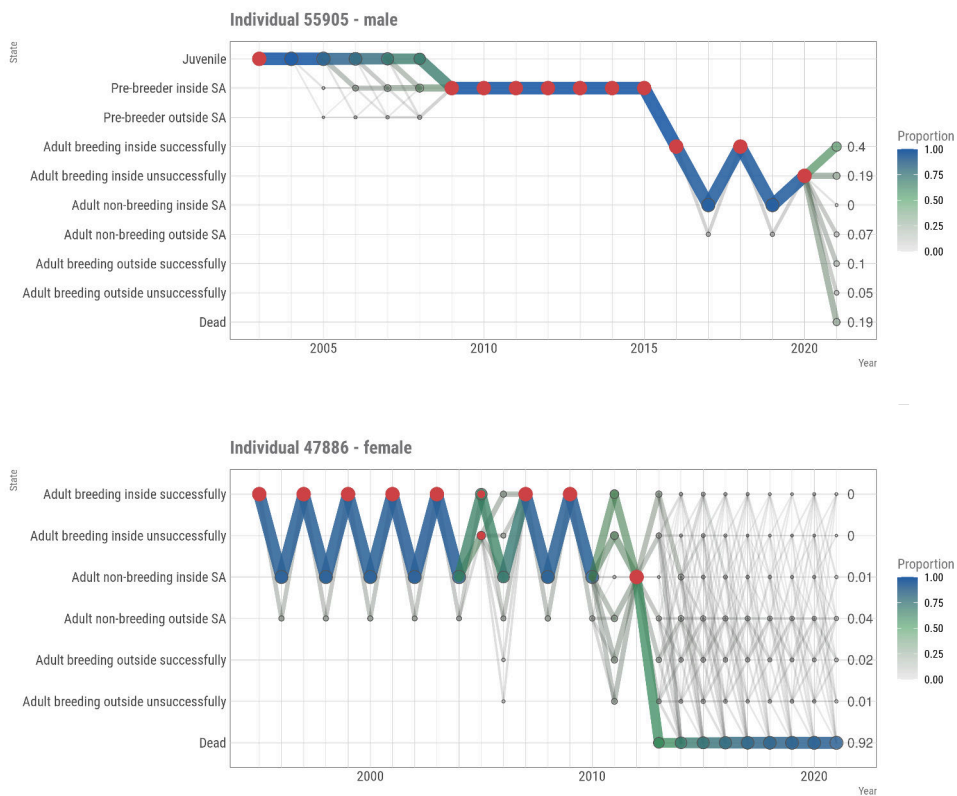


Figure 7: Examples of drawing the latent state of individuals from their observed state. Red dots represent the latent states that are possible given the observed state of an individual that was detected. The size and colour of segments indicate the probability of transition between two successive states. Numbers indicate the probability of each state in 2021, used to draw the initial population structure for population projections.

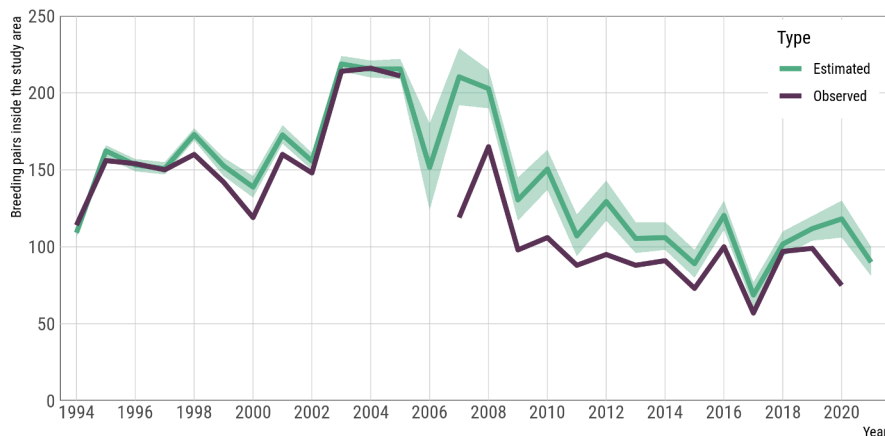
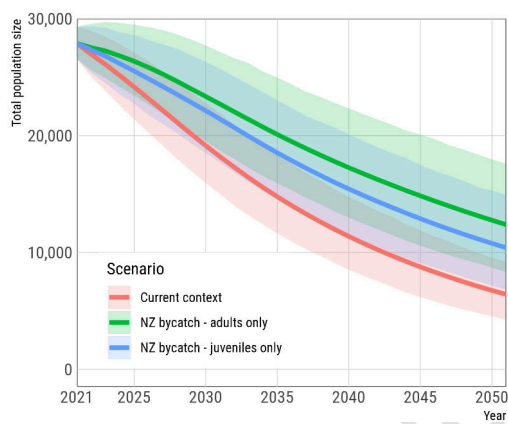


Figure 8: Comparison of the annual number of breeding pairs when recorded during field surveys (Observed) and when estimated from the model (Estimated). Lines indicate the mean, shading the 95% credible interval for the estimate.

(a) Total population



(b) Total breeding pairs

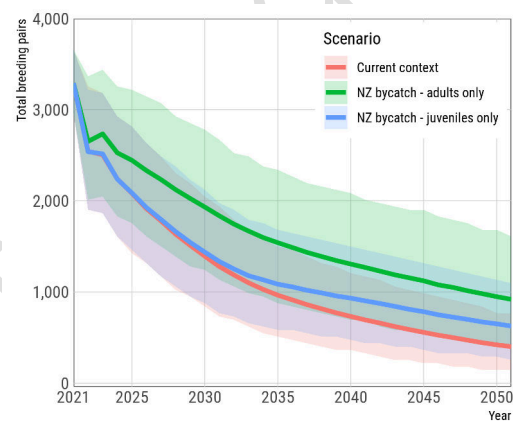


Figure 9: Simulation of the population size (a) and of the number of annual breeding pairs (b) of Antipodean albatross over the next 30 years on Antipodes Island. The simulation is based on the demographic parameters estimated in the model, only keeping the time-varying values between 2008 and 2020. The mean and 95% credible interval are shown.

374 4. DISCUSSION

The aim of this project was to provide an online simulation tool for predicting the
376 outcome of management strategies on the demography of Antipodean albatross. As for
any model, the accuracy of the prediction depends on the input field data, the complexity
378 of the factors affecting the demography, and the change over time in the threats to the
species.

380 Although movements in and out of the study area were included in the model, any
permanent emigration from the study area was more likely to be considered as local
382 mortality, and may underestimate annual survival rate. The area around the study
site has been visited regularly and sightings recorded there were used in the model to
384 estimate the rate of movements between areas. It is a relatively small area compared
with the rest of the island; some individuals may not be seen again once they relocate
386 permanently, making their emigration indistinguishable from death. Nevertheless, the
observations of the researchers when travelling across the island suggest that permanent
388 emigration by a significant number of individuals is unlikely (G. Elliott, pers. comm.).

The current model specification was designed to provide a basis for the simulations, and
390 compromises were made to balance realism and simplicity. For example, a number of
parameters were not dependent on years, such as the probability of breeding or the
392 survival rate of pre-breeders, and the model presented here may not be the closest
representation of reality. For this reason, the model results and the absolute projections
394 into the future need to be considered with caution. Nevertheless, it should be sufficient
to compare the relative impact of alternative management strategies. (Slight changes to
396 the model may be applied upon finalisation of this report.)

The recent increase in survival rates since 2018 may be a probabilistic coincidence, but
398 could also indicate an alleviation of the threats affecting females predominantly. For
example, fisheries may operate in different areas over time, or the areas where individuals
400 forage may also vary, resulting in a change in the overlap between the species and fishery
threats. The next few years of field data will inform whether this trend continues.

402 5. ACKNOWLEDGEMENTS

I am sincerely thankful to Graeme Elliott and Kath Walker for providing us with the
404 data they have tenaciously collected over decades, and for sharing their experience of
the species and field work.

406 The assistance of Philipp Neubauer in the development and fitting of the Bayesian model
in Stan was very useful and appreciated.

408 This work was funded by the Department of Conservation (through project BCBC2020-
09). I am grateful to Johannes Fischer and Igor Debski for the constructive talks during
410 the project.

Data preparation and statistical analyses were carried out using R (including the libraries
412 data.table and rstan) and Stan, writing scripts using Emacs, containerised using Docker,
and this document was produced using L^AT_EX. We are extremely grateful to the many
414 people who contribute to these key open source software projects and make them
available.

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454 APPENDIX A STAN MODEL CODE

```

1 functions{
450
3 matrix trans_probs(int nstates, real s_ad, real s_prebr, real s_juv,
451 real p_mv_out, real p_mv_in, int succ, vector p_breed,
5 real p_rec, real p_bead, real p_succ) {
460
7 /** TRANSITIONS and SURVIVAL */
461
9 // 1: adults breeding inside SA
462 // 2: adults breeding outside SA
11 // 3: adults non-breeding inside SA
463 // 4: adults non-breeding outside SA
13 // 5: pre-breeders inside SA
464 // 6: pre-breeders outside SA
15 // 7: juvs
470 // 8: deads
17
471 matrix[nstates, nstates] tmat;
19
472 /** ADULTS PREVIOUSLY BREEDING WITHIN STUDY AREA */
// re-breeding in SA (SA = study area)
473 tmat[1, 1] = succ == 2 ?
23 0 :
474 (succ == 1 ?
25 p_breed[1] * s_ad * (1-p_mv_out) :
480 (1-p_succ) * p_breed[1] * s_ad * (1-p_mv_out));
27 // re-breeding outside SA
481 tmat[1, 2] = succ == 2 ?
29 0 :
482 (succ == 1 ?
31 p_breed[1] * s_ad * p_mv_out :
483 (1-p_succ) * p_breed[1] * s_ad * p_mv_out);
33 // non-breeding in SA
484 tmat[1, 3] = succ == 2 ?
35 s_ad * (1-p_mv_out) :
490 (succ == 1 ?
37 (1-p_breed[1]) * s_ad * (1-p_mv_out) :
491 (1-p_succ) * (1-p_breed[1]) * s_ad * (1-p_mv_out) +
39 p_succ * s_ad * (1-p_mv_out));
492 // non-breeding outside SA
41 tmat[1, 4] = succ == 2 ?
493 s_ad * p_mv_out :
43 (succ == 1 ?
494 (1-p_breed[1]) * s_ad * p_mv_out :
45 (1-p_succ) * (1-p_breed[1]) * s_ad * p_mv_out +
500 p_succ * s_ad * p_mv_out);
47 tmat[1, 5] = 0; // pre-breeders inside SA
501 tmat[1, 6] = 0; // pre-breeders outside SA
49 tmat[1, 7] = 0; // juvs
502 tmat[1, 8] = 1-s_ad; // dead
51
503 /** ADULTS PREVIOUSLY BREEDING OUTSIDE STUDY AREA */
53 // re-breeding in SA (SA = study area)
504 tmat[2, 1] = succ == 2 ?
55 0 :
551 (succ == 1 ?
57 p_breed[1] * s_ad * p_mv_in :
552 (1-p_succ) * p_breed[1] * s_ad * p_mv_in);
59 // re-breeding outside SA
560 tmat[2, 2] = succ == 2 ?
61 0 :
561 (succ == 1 ?
63 p_breed[1] * s_ad * (1-p_mv_in) :
562 (1-p_succ) * p_breed[1] * s_ad * (1-p_mv_in));
65 // non-breeding in SA
563 tmat[2, 3] = succ == 2 ?
67 s_ad * p_mv_in :
564 (succ == 1 ?
69 (1-p_breed[1]) * s_ad * p_mv_in :
570 (1-p_succ) * (1-p_breed[1]) * s_ad * p_mv_in + p_succ * s_ad * p_mv_in);
71 // non-breeding outside SA
571 tmat[2, 4] = succ == 2 ?
73 s_ad * (1-p_mv_in) :
572 (succ == 1 ?
75 (1-p_breed[1]) * s_ad * (1-p_mv_in) :
573 (1-p_succ) * (1-p_breed[1]) * s_ad * (1-p_mv_in) + p_succ * s_ad * (1-p_mv_in));
77 tmat[2, 5] = 0; // pre-breeders inside SA
574 tmat[2, 6] = 0; // pre-breeders outside SA
79 tmat[2, 7] = 0; // juvs
580 tmat[2, 8] = 1-s_ad; // dead
81
581 /** ADULTS PREVIOUSLY NOT BREEDING WITHIN STUDY AREA */
83 tmat[3, 1] = p_breed[2] * s_ad * (1-p_mv_out); // breeding in SA (SA = study area)
582 tmat[3, 2] = p_breed[2] * s_ad * p_mv_out; // breeding outside SA
85 tmat[3, 3] = (1-p_breed[2]) * s_ad * (1-p_mv_out); // non-breeding in SA
583 tmat[3, 4] = (1-p_breed[2]) * s_ad * p_mv_out; // non-breeding outside SA
87 tmat[3, 5] = 0; // pre-breeders inside SA
584 tmat[3, 6] = 0; // pre-breeders outside SA
89 tmat[3, 7] = 0; // juvs

```

```

500 tmat[3, 8] = 1-s_ad; // dead
91
500 /** ADULTS PREVIOUSLY NOT BREEDING OUTSIDE THE STUDY AREA */
93 tmat[4, 1] = p_breed[2] * s_ad * p_mv_in; // breeding in SA (SA = study area)
500 tmat[4, 2] = p_breed[2] * s_ad * (1-p_mv_in); // breeding outside SA
95 tmat[4, 3] = (1-p_breed[2]) * s_ad * p_mv_in; // non-breeding in SA
500 tmat[4, 4] = (1-p_breed[2]) * s_ad * (1-p_mv_in); // non-breeding outside SA
97 tmat[4, 5] = 0; // pre-breeders inside SA
500 tmat[4, 6] = 0; // pre-breeders outside SA
99 tmat[4, 7] = 0; // juvs
500 tmat[4, 8] = 1-s_ad; // dead
101
500 /** PRE-BREEDERS INSIDE THE STUDY AREA */
103 tmat[5, 1] = s_prebr * p_bead * (1-p_mv_out); // breeding in SA (SA = study area)
500 tmat[5, 2] = s_prebr * p_bead * p_mv_out; // breeding outside SA
105 tmat[5, 3] = 0; // non-breeding in SA
500 tmat[5, 4] = 0; // non-breeding outside SA
107 tmat[5, 5] = s_prebr * (1-p_bead) * (1-p_mv_out); // pre-breeders inside SA
500 tmat[5, 6] = s_prebr * (1-p_bead) * p_mv_out; // pre-breeders outside SA
109 tmat[5, 7] = 0; // juvs
500 tmat[5, 8] = 1-s_prebr; // dead
111
500 /** PRE-BREEDERS OUTSIDE THE STUDY AREA */
113 tmat[6, 1] = s_prebr * p_bead * p_mv_in; // breeding in SA (SA = study area)
500 tmat[6, 2] = s_prebr * p_bead * (1-p_mv_in); // breeding outside SA
115 tmat[6, 3] = 0; // non-breeding in SA
500 tmat[6, 4] = 0; // non-breeding outside SA
117 tmat[6, 5] = s_prebr * (1-p_bead) * p_mv_in; // pre-breeders inside SA
500 tmat[6, 6] = s_prebr * (1-p_bead) * (1-p_mv_in); // pre-breeders outside SA
119 tmat[6, 7] = 0; // juvs
500 tmat[6, 8] = 1-s_prebr; // dead
121
500 /** JUVENILES */
123 tmat[7, 1] = 0; // breeding in SA (SA = study area)
500 tmat[7, 2] = 0; // breeding outside SA
125 tmat[7, 3] = 0; // non-breeding in SA
500 tmat[7, 4] = 0; // non-breeding outside SA
127 tmat[7, 5] = s_juv * p_rec * (1-p_mv_out); // pre-breeders inside SA
500 tmat[7, 6] = s_juv * p_rec * p_mv_out; // pre-breeders outside SA
129 tmat[7, 7] = s_juv * (1-p_rec); // juvs
500 tmat[7, 8] = 1-s_juv; // dead
131
500 /** DEADS */
133 tmat[8, 1] = 0; // breeding in SA (SA = study area)
500 tmat[8, 2] = 0; // breeding outside SA
135 tmat[8, 3] = 0; // non-breeding in SA
500 tmat[8, 4] = 0; // non-breeding outside SA
137 tmat[8, 5] = 0; // pre-breeders inside SA
500 tmat[8, 6] = 0; // pre-breeders outside SA
139 tmat[8, 7] = 0; // juvs
500 tmat[8, 8] = 1; // dead
141
500 return tmat;
143 }
145
600 matrix obs_probs(int n_obs_states, real[] p_obs, real p_detect_juv, real p_detect_dead,
147 real p_female, real p_succ, int succ, int no_visit) {
600
149 /** OBSERVED STATES */
600
151 // 1: adults breeding in SA
600 // 2: adults non-breeding in SA
153 // 3: adults outside SA
600 // 4: pre-breeders inside SA
155 // 5: pre-breeders outside SA
600 // 6: juvs
157 // 7: dead
600 // 8: not seen
159
600 matrix[n_obs_states, n_obs_states] pmat;
161
600 /** ADULTS BREEDING WITHIN STUDY AREA */
163 pmat[1, 1] = no_visit == 1 ? 0 : p_obs[1]; // ad breeding in SA (SA = study area)
600 pmat[1, 2] = 0; // ad non-breeding in SA
165 pmat[1, 3] = 0; // ad outside SA
600 pmat[1, 4] = 0; // pre-breeders inside SA
167 pmat[1, 5] = 0; // pre-breeders outside SA
600 pmat[1, 6] = 0; // juvs
169 pmat[1, 7] = 0; // dead
600 pmat[1, 8] = 1 - pmat[1, 1]; // not seen
171
600 /** ADULTS BREEDING OUTSIDE STUDY AREA */
173 pmat[2, 1] = 0; // ad breeding in SA (SA = study area)
600 pmat[2, 2] = 0; // ad non-breeding in SA
175 pmat[2, 3] = no_visit == 1 ? 0 : p_obs[5]; // ad outside SA
600 pmat[2, 4] = 0; // pre-breeders inside SA
177 pmat[2, 5] = 0; // pre-breeders outside SA
600 pmat[2, 6] = 0; // juvs
179 pmat[2, 7] = 0; // dead
600 pmat[2, 8] = 1 - pmat[2, 3]; // not seen

```

```

181
688 /** ADULTS NON-BREEDING INSIDE STUDY AREA */
183 pmat[3, 1] = 0; // ad breeding in SA (SA = study area)
684 pmat[3, 2] = no_visit == 1 ? // ad non-breeding in SA
185 0 :
686 (succ == 2 ?
187 p_obs[2] :
688 (succ == 1 ?
189 p_obs[3] :
690 p_succ * p_obs[2] + (1-p_succ) * p_obs[3]));
191 pmat[3, 3] = 0; // ad outside SA
692 pmat[3, 4] = 0; // pre-breeders inside SA
193 pmat[3, 5] = 0; // pre-breeders outside SA
694 pmat[3, 6] = 0; // juvs
195 pmat[3, 7] = 0; // dead
696 pmat[3, 8] = 1 - pmat[3, 2]; // not seen
197
698 /** ADULTS NON-BREEDING OUTSIDE STUDY AREA */
199 pmat[4, 1] = 0; // ad breeding in SA (SA = study area)
699 pmat[4, 2] = 0; // ad non-breeding in SA
201 pmat[4, 3] = no_visit == 1 ? 0 : p_obs[5]; // ad outside SA
699 pmat[4, 4] = 0; // pre-breeders inside SA
203 pmat[4, 5] = 0; // pre-breeders outside SA
699 pmat[4, 6] = 0; // juvs
205 pmat[4, 7] = 0; // dead
699 pmat[4, 8] = 1 - pmat[4, 3]; // not seen
207
698 /** PRE-BREEDERS INSIDE STUDY AREA */
209 pmat[5, 1] = 0; // ad breeding in SA (SA = study area)
699 pmat[5, 2] = 0; // ad non-breeding in SA
211 pmat[5, 3] = 0; // ad outside SA
699 pmat[5, 4] = no_visit == 1 ? 0 : p_obs[4]; // pre-breeders inside SA
213 pmat[5, 5] = 0; // pre-breeders outside SA
699 pmat[5, 6] = 0; // juvs
215 pmat[5, 7] = 0; // dead
699 pmat[5, 8] = 1 - pmat[5, 4]; // not seen
217
698 /** PRE-BREEDERS OUTSIDE STUDY AREA */
219 pmat[6, 1] = 0; // ad breeding in SA (SA = study area)
699 pmat[6, 2] = 0; // ad non-breeding in SA
221 pmat[6, 3] = 0; // ad outside SA
699 pmat[6, 4] = 0; // pre-breeders inside SA
223 pmat[6, 5] = no_visit == 1 ? 0 : p_obs[5]; // pre-breeders outside SA
699 pmat[6, 6] = 0; // juvs
225 pmat[6, 7] = 0; // dead
699 pmat[6, 8] = 1 - pmat[6, 5]; // not seen
227
698 /** JUVENILES */
229 pmat[7, 1] = 0; // ad breeding in SA (SA = study area)
699 pmat[7, 2] = 0; // ad non-breeding in SA
231 pmat[7, 3] = 0; // ad outside SA
699 pmat[7, 4] = 0; // pre-breeders inside SA
233 pmat[7, 5] = 0; // pre-breeders outside SA
699 pmat[7, 6] = no_visit == 1 ? 0 : p_detect_juv; // juvs
235 pmat[7, 7] = 0; // dead
699 pmat[7, 8] = 1 - pmat[7, 6]; // not seen
237
698 /** DEADS */
239 pmat[8, 1] = 0; // ad breeding in SA (SA = study area)
699 pmat[8, 2] = 0; // ad non-breeding in SA
241 pmat[8, 3] = 0; // ad outside SA
699 pmat[8, 4] = 0; // pre-breeders inside SA
243 pmat[8, 5] = 0; // pre-breeders outside SA
699 pmat[8, 6] = 0; // juvs
245 pmat[8, 7] = no_visit == 1 ? 0 : p_detect_dead; // dead
700 pmat[8, 8] = 1 - pmat[8, 7]; // not seen
247
208 return pmat;
249 }
250
251
252 real log_sum_one_indiv (int N_STATES, int sex, int[] age, int MAX_T, int first_cap, int last_cap,
253 int[] c_hist, real[,] s_ad, real s_prebr, real s_juv, real[] p_moveout,
254 real[] p_movein, int[] b_success, vector p_breed, vector p_recruit,
255 vector p_beadult, real[] p_success, int N_STATES_P, real[,] p_obs,
256 real p_detect_juv, real p_detect_dead, real p_female, int[] NO_VISIT,
257 int first_state) {
258
259 matrix[N_STATES, N_STATES] tmat;
260 matrix[N_STATES_P, N_STATES_P] pmat;
261 vector[N_STATES] pz[MAX_T];
262 real temp[N_STATES];
263 real lsum;
264
265 for (j in 1:N_STATES) {
266 pz[first_cap, j] = (j == first_state);
267 }
268 for (t in (first_cap+1):last_cap) {
269 tmat = trans_probs(N_STATES, s_ad[sex+1, t-1], s_prebr, s_juv,
270 p_moveout[sex+1], p_movein[sex+1], b_success[t-1],
271 p_breed, p_recruit[age[t]], p_beadult[age[t]],

```

```

220     p_success[t-1]);
273     pmat = obs_probs(N_STATES_P, p_obs[t-1], p_detect_juv, p_detect_dead, p_female,
228     p_success[t-1], b_success[t-1], NO_VISIT[t]);
275     for (i in 1:N_STATES) {
230       for (j in 1:N_STATES) {
277         temp[j] = pz[t-1, j] * tmat[j, i] * pmat[i, c_hist[t]];
238       }
279       pz[t, i] = sum(temp);
280     }
281   }
282 }
283 lsum = log(sum(pz[last_cap]));
284
285 return lsum;
286
287 }
288
289
290 real calc_log_sum_multi (int[] INDS, int start, int end, int N_STATES, int[] SEX, int[,] AGE,
291 int MAX_T, int[] FIRST_CAP, int[] LAST_CAP, int[,] C_HIST,
292 real[,] s_ad, real s_prebr, real s_juv,
293 real[] p_moveout, real[] p_movein, int[,] B_SUCCESS, vector p_breed,
294 vector p_recruit, vector p_beadult, real[] p_success,
295 int N_STATES_P, real[,] p_obs, real p_detect_juv, real p_detect_dead, real p_female,
296 int[] NO_VISIT, int[] FIRST_STATE) {
297
298   real lsum;
299
300   lsum = 0.0;
301   for (ind in start:end) {
302     lsum += log_sum_one_indiv(N_STATES, SEX[ind], AGE[ind], MAX_T, FIRST_CAP[ind], LAST_CAP[ind], C_HIST[ind],
303     s_ad, s_prebr, s_juv, p_moveout, p_movein,
304     B_SUCCESS[ind], p_breed,
305     p_recruit, p_beadult, p_success,
306     N_STATES_P, p_obs, p_detect_juv, p_detect_dead, p_female,
307     NO_VISIT, FIRST_STATE[ind]); //, ind);
308   }
309
310   return lsum;
311 }
312
313 }
314
315
316 data {
317
318   int<lower=1> N_INDS;
319   int<lower=1> INDS [N_INDS];
320   int<lower=1> FIRST_STATE [N_INDS];
321
322   int<lower=0, upper=2> SEX [N_INDS];
323   int<lower=1> N_SEXED;
324   int<lower=0, upper=1> IS_FEMALE [N_SEXED];
325
326   int<lower=1> N_NESTS;
327   int<lower=0, upper=1> NEST_SUCCESS [N_NESTS];
328   int<lower=1> NEST_YEAR [N_NESTS];
329
330   int<lower=1> FIRST_CAP [N_INDS];
331   int<lower=1> LAST_CAP [N_INDS];
332
333   int<lower=1> MAX_T;
334   int<lower=1> MAX_AGE;
335   int<lower=1> AGE[N_INDS, MAX_T];
336
337   int<lower=1, upper=MAX_AGE> MIN_R_AGE;
338   int<lower=MIN_R_AGE, upper=MAX_AGE> MAX_R_AGE;
339   int<lower=MIN_R_AGE, upper=MAX_AGE> MIN_B_AGE;
340   int<lower=MIN_B_AGE, upper=MAX_AGE> MIN_B_AGE2;
341
342   int<lower=0, upper=1> NO_VISIT[MAX_T];
343
344   int<lower=1> N_STATES;
345   int<lower=1> N_STATES_P;
346   int<lower=1> N_PDETECTS;
347   int<lower=1, upper=N_STATES_P> C_HIST[N_INDS, MAX_T];
348
349   int<lower=0, upper=2> B_SUCCESS[N_INDS, MAX_T];
350 }
351
352
353
354 transformed data {
355   int<lower=1> grainsize=1;
356 }
357
358
359 parameters{
360
361   real<lower=0, upper=1> p_female;
362

```

```

363 real<lower=0, upper=1> p_rec [MAX_R_AGE - MIN_R_AGE + 1];
364
365 real<lower=0, upper=1> p_br [MIN_B_AGE2 - MIN_B_AGE];
366 real<lower=0, upper=1> p_br_post;
367
368 vector<lower=0, upper=1>[2] p_breed; // 1: previously unsuccessful breeders; 2: other non-breeders
369
370 real<lower=0> sigma_re_bsucc;
371 real bsucc_lg_re [MAX_T];
372 real bsucc_lg_mean;
373
374 real<lower=0, upper=1> s_prebr;
375 real<lower=0, upper=1> s_juv;
376
377 /* Random effect on recruitment */
378 real<lower=0> sigma_re_rec;
379 real rec_lg_re [MAX_R_AGE - MIN_R_AGE + 1];
380 real rec_lg_mean;
381
382 /* Random effect on becoming adult */
383 real<lower=0> sigma_re_bead;
384 real bead_lg_re [MIN_B_AGE2 - MIN_B_AGE];
385 real bead_lg_mean;
386
387 /* Random effect on adult survival */
388 real<lower=0> sigma_re_ad_s;
389 real surv_ad_lg_re [2, MAX_T-1];
390 real surv_ad_lg_mean [2];
391
392 /* Random effect on detectability */
393 real<lower=0> sigma_re_p;
394 real p_detect_lg_re [MAX_T-1];
395 real p_detect_lg_mean [N_PDETECTS];
396 real<lower=0, upper=1> p_detect_juv;
397 real<lower=0, upper=1> p_detect_dead;
398
399 real<lower=0, upper=1> p_leave[2];
400 real<lower=0, upper=1> p_back[2];
401
402 }
403
404 transformed parameters {
405
406 vector<lower=0, upper=1>[MAX_AGE] p_recruit;
407 vector<lower=0, upper=1>[MAX_AGE] p_beadult;
408
409 real<lower=0, upper=1> s_adult [2, MAX_T-1];
410 real<lower=0, upper=1> s_ad[3, MAX_T-1];
411
412 real<lower=0, upper=1> p_detect [N_PDETECTS, MAX_T-1]; // 1: breeding ad (inside sa); 2: non-breeding ad previously
413 successful (inside sa); 3: other non-breeders (inside sa); 4: prebr inside SA; 5: ad or prebr outside SA
414
415 real<lower=0, upper=1> p_success [MAX_T];
416
417 real<lower=0, upper=1> p_moveout [3];
418 real<lower=0, upper=1> p_movein [3];
419
420 real<lower=0, upper=1> p_obs [MAX_T-1, N_PDETECTS];
421
422 /* Juvs becoming pre-breeders (recruitment to the colony) */
423 for (a in 1:(MIN_R_AGE-1)) {
424   p_recruit[a] = 0;
425 }
426 for (a in MIN_R_AGE:MAX_R_AGE) {
427   p_recruit[a] = inv_logit(rec_lg_mean + rec_lg_re[a - MIN_R_AGE + 1] * sigma_re_rec);
428 }
429 for (a in (MAX_R_AGE+1):MAX_AGE) {
430   p_recruit[a] = 1;
431 }
432
433 /* Pre-Breeders becoming adults (start breeding) */
434 for (a in 1:(MIN_B_AGE-1)) {
435   p_beadult[a] = 0;
436 }
437 for (a in MIN_B_AGE:(MIN_B_AGE2-1)) {
438   p_beadult[a] = inv_logit(bead_lg_mean + bead_lg_re[a - MIN_B_AGE + 1] * sigma_re_bead);
439 }
440 for (a in MIN_B_AGE2:MAX_AGE) {
441   p_beadult[a] = p_br_post;
442 }
443
444 for (t in 1:(MAX_T-1)) {
445   for (s in 1:N_PDETECTS) {
446     p_detect[s, t] = inv_logit(p_detect_lg_mean[s] + p_detect_lg_re[t] * sigma_re_p);
447   }
448   for (sex in 1:2) {
449     s_adult[sex, t] = inv_logit(surv_ad_lg_mean[sex] + surv_ad_lg_re[sex, t] * sigma_re_ad_s);
450   }
451 }
452 }




































```



```
544
1000 /* Likelihood */
546 target += reduce_sum(calc_log_sum_multi, INDS, grainsize, N_STATES, SEX, AGE, MAX_T,
1002 FIRST_CAP, LAST_CAP, C_HIST,
548 s_ad, s_prebr, s_juv, p_moveout, p_movein,
1000 B_SUCCESS, p_breed,
550 p_recruit, p_beadult, p_success,
1006 N_STATES_P, p_obs, p_detect_juv, p_detect_dead, p_female,
552 NO_VISIT, FIRST_STATE);
1008
554 }
```

DRAFT - Not to be quoted

Table B-1: Annual survival rate of adults by year and sex, and of pre-breeders and juveniles. Shown are the mean, 95% credible interval (c.i.), and the MCMC trace of the parameter.

Year	Females			Males		
	Mean	95% c.i.	Trace	Mean	95% c.i.	Trace
1994	0.961	0.924 – 0.987		0.959	0.919 – 0.986	
1995	0.970	0.944 – 0.989		0.963	0.933 – 0.986	
1996	0.951	0.917 – 0.977		0.936	0.901 – 0.966	
1997	0.950	0.916 – 0.975		0.950	0.916 – 0.978	
1998	0.924	0.884 – 0.959		0.936	0.900 – 0.965	
1999	0.923	0.881 – 0.959		0.917	0.879 – 0.949	
2000	0.941	0.901 – 0.972		0.941	0.908 – 0.968	
2001	0.935	0.898 – 0.965		0.951	0.924 – 0.975	
2002	0.951	0.925 – 0.973		0.949	0.921 – 0.973	
2003	0.975	0.954 – 0.990		0.965	0.942 – 0.983	
2004	0.941	0.906 – 0.971		0.936	0.907 – 0.960	
2005	0.898	0.806 – 0.969		0.941	0.889 – 0.980	
2006	0.878	0.789 – 0.963		0.933	0.883 – 0.978	
2007	0.883	0.822 – 0.936		0.897	0.853 – 0.935	
2008	0.845	0.776 – 0.911		0.920	0.879 – 0.954	
2009	0.894	0.832 – 0.947		0.914	0.874 – 0.949	
2010	0.906	0.840 – 0.960		0.926	0.889 – 0.958	
2011	0.841	0.770 – 0.906		0.937	0.901 – 0.969	
2012	0.829	0.760 – 0.891		0.911	0.872 – 0.943	
2013	0.821	0.752 – 0.883		0.936	0.902 – 0.963	
2014	0.929	0.876 – 0.970		0.950	0.918 – 0.977	
2015	0.848	0.781 – 0.909		0.914	0.873 – 0.949	
2016	0.871	0.803 – 0.930		0.913	0.871 – 0.949	
2017	0.937	0.884 – 0.977		0.924	0.879 – 0.959	
2018	0.901	0.832 – 0.956		0.916	0.871 – 0.952	
2019	0.908	0.839 – 0.963		0.938	0.894 – 0.972	
2020	0.929	0.861 – 0.976		0.971	0.943 – 0.991	

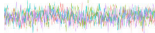

Age class	Mean	95% c.i.	Trace
Pre-breeders	0.922	0.913 – 0.931	
Juveniles	0.879	0.869 – 0.888	

Table B-2: Probabilities of successful breeding by year. Shown are the mean, 95% credible interval (c.i.), and the MCMC trace of the parameter.

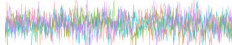


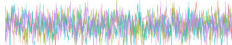







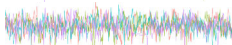






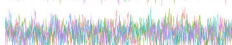






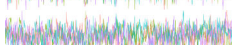

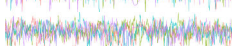
Parameter	Year	Mean	95% c.i.	Trace
P(successful breeding)	1994	0.74	0.67 – 0.81	
	1995	0.74	0.67 – 0.79	
	1996	0.75	0.68 – 0.81	
	1997	0.77	0.71 – 0.83	
	1998	0.73	0.66 – 0.79	
	1999	0.64	0.56 – 0.71	
	2000	0.75	0.67 – 0.81	
	2001	0.75	0.69 – 0.81	
	2002	0.67	0.60 – 0.74	
	2003	0.72	0.66 – 0.77	
	2004	0.71	0.65 – 0.77	
	2005	0.69	0.53 – 0.83	
	2006	0.67	0.48 – 0.81	
	2007	0.60	0.52 – 0.67	
	2008	0.67	0.60 – 0.73	
	2009	0.56	0.47 – 0.65	
	2010	0.67	0.58 – 0.74	
	2011	0.57	0.48 – 0.66	
	2012	0.58	0.49 – 0.67	
	2013	0.61	0.51 – 0.69	
	2014	0.70	0.62 – 0.77	
2015	0.57	0.47 – 0.66		
2016	0.66	0.57 – 0.74		
2017	0.75	0.65 – 0.83		
2018	0.65	0.57 – 0.73		
2019	0.61	0.53 – 0.69		
2020	0.62	0.52 – 0.71		
2021	0.67	0.48 – 0.81		

Table B-3: Probabilities of returning to the colony and to breed for the first time, as function of age. Shown are the mean, 95% credible interval (c.i.), and the MCMC trace of the parameter.

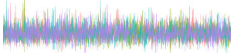
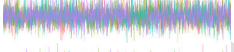
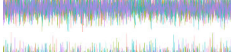
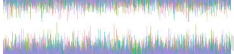




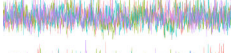
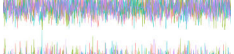
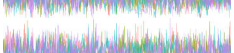

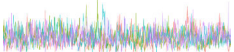
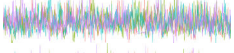

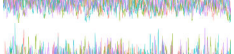




Parameter	Age	Mean	95% c.i.	Trace
P(return to colony)	3	0.03	0.02 – 0.04	
	4	0.11	0.09 – 0.14	
	5	0.10	0.07 – 0.14	
	6	0.29	0.24 – 0.34	
	7	0.33	0.25 – 0.42	
	8	0.69	0.56 – 0.83	
P(breed for first time)	7	0.02	0.01 – 0.04	
	8	0.06	0.04 – 0.09	
	9	0.07	0.05 – 0.10	
	10	0.06	0.04 – 0.09	
	11	0.10	0.07 – 0.14	
	12	0.13	0.09 – 0.18	
	13	0.12	0.08 – 0.17	
	14	0.13	0.09 – 0.18	
	15	0.07	0.04 – 0.12	
	16	0.13	0.08 – 0.19	
	17	0.10	0.05 – 0.16	
	18	0.09	0.04 – 0.15	
	19	0.12	0.06 – 0.20	
	20	0.05	0.02 – 0.10	

Table B-4: Probabilities of adults breeding, that an individual is female, that a bird inside the study area move outside it, and probability that a bird outside the study area returns inside. Shown are the mean, 95% credible interval (c.i.), and the MCMC trace of the parameter.

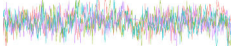
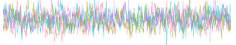
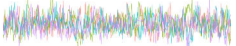

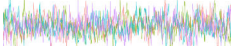
Parameter	Category	Mean	95% c.i.	Trace
P(breeding)	Previously unsuccessful breeders	0.70	0.69 – 0.72	
	Other non-breeders	0.64	0.63 – 0.65	
P(female)		0.51	0.49 – 0.53	
P(leave the study area)	Female	0.09	0.08 – 0.10	
	Male	0.04	0.04 – 0.05	
P(return to the study area)	Female	0.18	0.15 – 0.20	
	Male	0.25	0.22 – 0.29	

Table B-5: Detection probabilities: annual averages, year effect, and interannual variability, as well as the time-invariant detection probabilities of juveniles and dead birds. Shown are the mean, 95% credible interval (c.i.), and the MCMC trace of the parameter.

Parameter	Category	Mean	95% c.i.	Trace
P(detection) - overall	Breeding adult (inside SA)	0.864	0.816 – 0.900	
	Non-breeding adult (inside SA)	0.052	0.036 – 0.072	
	Other non-breeders (inside SA)	0.997	0.992 – 1.000	
	Pre-breeders (inside SA)	0.661	0.575 – 0.736	
	Adults and pre-breeders outside SA	0.180	0.132 – 0.234	
Year effect (logit scale)	1995	-1.388	-2.214 – -0.679	
	1996	1.014	0.471 – 1.605	
	1997	1.529	0.915 – 2.168	
	1998	0.711	0.189 – 1.294	
	1999	0.596	0.058 – 1.152	
	2000	0.238	-0.282 – 0.766	
	2001	0.475	-0.028 – 1.007	
	2002	1.052	0.494 – 1.657	
	2003	1.656	1.044 – 2.305	
	2004	1.475	0.901 – 2.078	
	2005	1.139	0.612 – 1.692	
	2006	0.017	-1.792 – 1.849	
	2007	-1.887	-2.654 – -1.229	
	2008	-0.491	-0.973 – -0.039	
	2009	-0.492	-0.982 – -0.034	
	2010	-0.394	-0.859 – 0.050	
	2011	-1.019	-1.593 – -0.507	
	2012	-0.464	-0.955 – 0.027	
	2013	0.004	-0.457 – 0.462	
	2014	0.329	-0.156 – 0.816	
	2015	-0.383	-0.885 – 0.124	
2016	-0.395	-0.931 – 0.100		
2017	-0.079	-0.585 – 0.424		
2018	0.229	-0.302 – 0.726		
2019	-0.021	-0.527 – 0.484		
2020	-1.763	-2.506 – -1.113		
2021	0.018	-0.536 – 0.565		
Inter-annual variability		0.858	0.638 – 1.169	

Parameter	Mean	95% c.i.	Trace
P(detection) - Juveniles	0.0002	0.0000 – 0.0007	
P(detection) - Dead birds	0.0008	0.0005 – 0.0012	